

1 **Modelling the migration of contaminants through variably**  
2 **saturated dual-porosity, dual-permeability chalk**

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## 1 **Abstract**

2 In the Hesbaye region in Belgium, tracer tests performed in variably saturated fissured chalk  
3 rocks presented very contrasting results in terms of transit times, according to artificially  
4 controlled water recharge conditions prevailing during the experiments. Under intense  
5 recharge conditions, tracers migrated across the partially or fully saturated fissure network, at  
6 high velocity in accordance with the high hydraulic conductivity and low effective porosity  
7 (fracture porosity). At the same time, a portion of the tracer was temporarily retarded in the  
8 almost immobile water located in the matrix. Under natural infiltration conditions, the fissure  
9 network remained inactive. Tracers migrated downward through the matrix, at low velocity in  
10 relation with the low hydraulic conductivity and the large porosity of the matrix. Based on  
11 these observations, Brouyère et al. (2004a) proposed a conceptual model in order to explain  
12 the migration of solutes in variably saturated, dual-porosity, dual-permeability chalk. Here,  
13 mathematical and numerical modelling of tracer and contaminant migration in variably  
14 saturated fissured chalk is presented, considering the aforementioned conceptual model. A  
15 new mathematical formulation is proposed to represent the unsaturated properties of the  
16 fissured chalk in a more dynamic and appropriate way. At the same time, the rock water  
17 content is partitioned between mobile and immobile water phases, as a function of the water  
18 saturation of the chalk rock. The groundwater flow and contaminant transport in the variably  
19 saturated chalk is solved using the control volume finite element method. Modelling the field  
20 tracer experiments performed in the variably saturated chalk shows the adequacy and  
21 usefulness of the new conceptual, mathematical and numerical model.

22 **Keywords:** dual-porosity, dual-permeability, fissured chalk, transport model, unsaturated  
23 zone, retention curve, relative hydraulic conductivity

## 1 **Introduction**

2 Recently, Brouyère et al. (2004a) have presented the results of an experimental study  
3 performed in the Hesbaye region in Belgium in order to characterize and to quantify  
4 hydrodynamic and hydrodispersive processes governing the downward migration of solute  
5 contaminants (e.g., nitrates) across the unsaturated zone overlying a fissured chalk aquifer.  
6 One of the most significant observations drawn from these experiments is the high contrast in  
7 terms of tracer transit times across the unsaturated chalk depending on the application of an  
8 artificial water recharge (forced gradient conditions) in the injection well or not (natural  
9 infiltration conditions). Tracer transit times across the unsaturated zone varied from a few  
10 hours when forced gradient conditions prevailed to almost one year under natural infiltration  
11 conditions.

12 The tracer test results can be explained by the dual-porosity, dual-permeability of the chalk. In  
13 the fissured chalk, groundwater flow and transport conditions can be highly variable  
14 according to the degree of water saturation of the rock. Under normal recharge conditions,  
15 fissures remain empty in the unsaturated zone of the chalk and a slow regime of flow and  
16 contaminant transport is active in the matrix characterised by a low hydraulic conductivity  
17 and a high porosity. In the saturated zone, or under intensive recharge conditions in the  
18 unsaturated zone, water and contaminants migrate at high velocity along the partially or fully  
19 saturated fissures controlling the hydraulic conductivity of the rock mass. The effective  
20 porosity is low, representing the contribution of fissures to the total porosity of the chalk. At  
21 the same time, transported contaminants are subject to retardation in the matrix (matrix  
22 diffusion) where the water can be considered as immobile compared to the water moving in  
23 the fissure network. The transport mechanism across the fissured chalk is schematised in  
24 Figure 1.

1 This conceptualisation of the dynamic behaviour of chalk hydraulic and hydrodispersive  
2 properties explained, from a conceptual point of view, the tracer test results (Brouyère et al.  
3 2004a). However, a transcription into a mathematical model was required. Below, the  
4 mathematical model developed in order to represent more accurately the variations in  
5 hydraulic conductivity and effective porosity of the chalk with respect to infiltration  
6 conditions and degree of saturation of the chalk is presented. The numerical solution of  
7 groundwater flow and contaminant transport in the variably saturated chalk, using the three-  
8 dimensional finite element simulator SUFT3D is described. Modelling results of the field  
9 tracer experiments performed in the chalk, used to illustrate the developments, support the  
10 conceptual model and demonstrate that the mathematical model is adequate to explain the  
11 tracer experiments.

## 12 **1 Modelling the hydraulic and hydrodispersive behaviour of the variably** 13 **saturated chalk**

14 Many conceptual and mathematical models were proposed to simulate the hydrodynamic  
15 behaviour of structured media characterised by preferential flow paths such as fissured rocks  
16 or aggregate soils (e.g., Bai et al. 1993, Pruess et al. 1999, Berkowitz 2002). Most advanced  
17 concepts rely on a distinct modelling of the preferential flows paths and the matrix (e.g., Gwo  
18 1992, Dykhuizen 1987, 1990, VanderKwaak 1999). Such approaches are relatively  
19 complicated and they require a large number of parameters that are often not available for  
20 characterizing both hydraulic networks and their interactions. Consequently, they are  
21 frequently not suitable for practical (field scale) applications. Another solution is to consider  
22 the medium as a single continuum (e.g., Peters and Klavetter 1988, Berkowitz et al. 1988,  
23 Finsterle 2000) by assuming a pressure equilibrium between the matrix and the fissures. In  
24 that case, the mathematical relationships used to model the hydraulic behaviour of the rock

1 mass under variably saturated conditions have to be adapted to account for the dynamic  
2 evolution of flow conditions according to the degree of saturation of the rock. Two  
3 characteristic curves need to be defined: the retention curve  $\theta(h)$  linking the water content  $\theta$   
4 (dimensionless) to the applied suction head  $h$  (L), and the hydraulic conductivity curve  $k_r(\theta)$   
5 relating the evolution of the relative hydraulic conductivity  $k_r$  (dimensionless) with the water  
6 content  $\theta$ . Classical mathematical relationships used to model unsaturated properties (e.g.  
7 van Genuchten 1980, Mualem 1976) are not directly suited as they do not allow an accurate  
8 representation of the unsaturated behaviour of the underground medium close to saturation.  
9 These relationships implicitly rely on a uni-modal distribution of pore dimensions (Fredlung  
10 and Xing 1994). In structured media, such as in fissured chalk, this distribution is at least bi-  
11 modal, often multi-modal (Price et al. 1993, Younger and Elliot 1995). Several relationships,  
12 based on bi- or multi-modal distributions of pore space, were proposed in the literature for  
13 modelling the unsaturated properties of structured media (e.g., Smettem and Kirby 1990,  
14 Othmer et al. 1991, Ross and Smettem 1993, Durner 1994, Fredlung and Xing 1994, Fredlung  
15 et al. 1994, Mallants et al. 1997). However, these models, usually developed for representing  
16 the retention curve, do not allow an analytical derivation of the relationship between the  
17 hydraulic conductivity and the water content from the retention curve. Furthermore, they are  
18 usually not continuously derivable, a condition that is not essential but interesting if  
19 groundwater flow simulations require the computation of derivatives of the unsaturated  
20 characteristic curves, such as Newton linearization (e.g., Paniconi et al. 1991, Paniconi and  
21 Putti 1994). The new mathematical relationships proposed here after to model the unsaturated  
22 characteristic curves of the chalk overcome these problems and they still offer a relatively  
23 large flexibility for modelling the unsaturated properties of structured media.

## 1 **1.1 Retention curve**

2 For modelling the unsaturated properties of fissures, a model such as that of Wang and  
3 Narasimhan (1985) could probably be very suited. However, it requires a detailed knowledge  
4 of the morphology and distribution of fissures, which is not available here. The retention  
5 curve of the chalk is modelled here using the combination of two van Genuchten  
6 relationships, one defined for the matrix, one for the fissures.

7 Up to this point, the model used for the retention curve is similar to that of Ross and Smettem  
8 (1993) or Durner (1994). In contrast to these models that just sum the two retention curves to  
9 build the global functionality describing the retention curve of the structured medium, the  
10 approach proposed here forces a continuously derivable transition between the relations that  
11 describe the matrix and the fissure component respectively (Figure 2). Accordingly, it is  
12 possible to derive analytically the hydraulic conductivity curve from the retention curve  
13 across the whole range of water contents.

14 The relationship used to model the retention curve associated with the matrix is (Figure 2,  
15 light grey curve):

$$16 \quad \Theta_M = \frac{\theta - \theta_r}{\theta_{s,M} - \theta_r} = \left[ 1 + (\alpha_M h)^{n_M} \right]^{-m_M} \quad (1)$$

17 where  $\theta_r$  is the residual water content of the matrix;  $\theta_{s,M}$  (dimensionless), to be considered  
18 just as a fitting parameter, is the ‘equivalent’ saturated water content of the matrix, close but  
19 different from the matrix total porosity  $n_m$  (dimensionless);  $\alpha_M$  ( $L^{-1}$ ),  $n_M$  (dimensionless) and  
20  $m_M$  (dimensionless) are the van Genuchten parameters used to fit the portion of the retention  
21 curve associated with the matrix;  $\Theta_M$  (dimensionless) is the relative saturation of the matrix.

22 The relationship used to model the retention curve associated with the fissure component of  
23 the chalk is (Figure 2, dark grey curve):

$$\Theta_F = \frac{\theta - \theta_{r,F}}{\theta_s - \theta_{r,F}} = \left[ 1 + (\alpha_F h)^{n_F} \right]^{-m_F} \quad (2)$$

where  $\theta_{r,F}$  (dimensionless), to be considered just as a fitting parameter, is the ‘equivalent’ residual water content of the fissure;  $\theta_s$  (dimensionless) is the saturated water content of the chalk rock;  $\alpha_F$  ( $L^{-1}$ ),  $n_F$  (dimensionless) and  $m_F$  (dimensionless) are the van Genuchten parameters used to fit the portion of the retention curve associated with the fissures;  $\Theta_F$  (dimensionless) is the relative saturation of the fissures.

At the point  $(\theta_J, h_J)$  common to the matrix and fissure retention curves (Figure 2), continuity conditions can be expressed as follows:

$$\text{-continuity of the retention curve: } (\theta_J, h_J)_M = (\theta_J, h_J)_F \quad (3a)$$

$$\text{-continuity of the first derivative of the retention curve: } \left( \frac{\partial \theta}{\partial h} \right)_{J,M} = \left( \frac{\partial \theta}{\partial h} \right)_{J,F} \quad (3b)$$

This comes to solving the following equation system in terms of  $\theta_{r,F}$  and  $n_F$  (Brouyère 2001):

$$\chi(\theta_{r,F}, n_F) = \theta_{r,F} - \frac{\theta_r + (\theta_{s,M} - \theta_r) \left( 1 + (\alpha_M h_J)^{n_M} \right)^{-m_M} + \theta_s \left( 1 + (\alpha_F h_J)^{n_F} \right)^{-m_F}}{\left( 1 + \left( 1 + (\alpha_F h_J)^{n_F} \right)^{-m_F} \right)} = 0 \quad (4a)$$

$$\gamma(\theta_{r,F}, n_F) = - \frac{\alpha_M m_M (\theta_{s,M} - \theta_r) \Theta_{J,M}^{1/m_M} \left( 1 - \Theta_{J,M}^{1/m_M} \right)^{m_M}}{1 - m_M} + \frac{\alpha_F m_F (\theta_s - \theta_{r,F}) \Theta_{J,F}^{1/m_F} \left( 1 - \Theta_{J,F}^{1/m_F} \right)^{m_F}}{1 - m_F} = 0 \quad (4b)$$

In practice, the adjustment of the retention curve representative of the chalk rock is performed as follows. First, the porosities associated with the matrix and the fissures respectively are defined, the total of these two values being set equal to the saturated water content  $\theta_s$  of the rock. The van Genuchten parameters relative to the matrix ( $\alpha_M$ ,  $n_M$  and  $\theta_{s,M}$ ) are obtained by fitting Equation (1) on retention curves measured on chalk matrix samples (see Brouyère et

1 al. 2004a). An estimation has to be provided for the capillary rise ( $\psi_F = 1/\alpha_F$ ) in the fissure  
 2 network. Estimates of the capillary rise in the fractures can be found in the literature (e.g.,  
 3 Price et al. 1993). The parameters  $n_F$  and  $\theta_{r,F}$  are determined by solving the equation system  
 4 (4a) and (4b) in order to meet the conditions (3a) and (3b).

## 5 **1.2 Relative hydraulic conductivity curve**

6 In order to derive the relationship between the hydraulic conductivity and the water content  
 7 analytically from the retention curve, the model of Mualem (1976) is considered. This model  
 8 was used already for the estimation of the evolution of hydraulic conductivity in fissured  
 9 rocks (Peters and Klavetter 1988). The fundamental relationship of Mualem's model is:

$$10 \quad k_r(\theta) = S_e^P \left[ \frac{f(\theta)}{f(\theta_s)} \right]^2 = S_e^P \left[ \frac{\int_0^\theta \frac{d\theta}{h}}{\int_0^{\theta_s} \frac{d\theta}{h}} \right]^2 \quad (5)$$

11 In Equation (5),  $P$  is a parameter for which an optimal value of 0.5 was proposed and

$$12 \quad S_e = \frac{\theta - \theta_r}{\theta_s - \theta_r}.$$

13 When the matrix is partially saturated and the fissures are completely desaturated,  
 14 mathematical integration is performed along the retention curve describing the matrix  
 15 (Equation 1):

$$16 \quad f_M(\theta) = \int_{\theta_r}^{\theta} \frac{d\theta}{h} = \alpha_M (\theta_{s,M} - \theta_r) \left( 1 - \left( 1 - \Theta_M^{1/m_M} \right)^{m_M} \right) \quad (6a)$$

17 For a fully saturated matrix and partially saturated fissures, integration starts from  $\theta = \theta_j$  and  
 18 it is performed along the retention curve describing the fissures (Equation 2):

$$19 \quad f(\theta) = f_M(\theta_j) + f_F(\theta) \quad (6b)$$



1 with  $f_F(\theta) = \int_{\theta_J}^{\theta} \frac{d\theta}{h} = \alpha_F (\theta_s - \theta_{r,F}) \left[ \left( 1 - \Theta_{J,F}^{1/m_F} \right)^{m_F} - \left( 1 - \Theta_F^{1/m_F} \right)^{m_F} \right]$  (6c)

2  $\Theta_{J,F} = \frac{\theta_J - \theta_{r,F}}{\theta_s - \theta_{r,F}}$

3 The evaluation of Equation (6c) in  $\theta = \theta_s$  gives:

4  $f(\theta_s) = f_M(\theta_J) + f_F(\theta_s)$  (6d)

5 Introducing Equations (6b) and (6d) into Equation (5) provides the functionality to describe  
6 the evolution of the relative hydraulic conductivity with the water content  $k_r(\theta)$ . For  
7 completeness, Annex 1 provides the mathematical expressions for the derivatives  $d\theta/dh$  and  
8  $dk_r/dh$ , needed for numerical solution of the Richards equation using Newton-Raphson  
9 linearization.

10 Figure 3 shows the characteristic curves obtained using mean van Genuchten parameters  
11 estimated by fitting the retention curves measured on the chalk matrix samples and using  
12 literature values for the fissure characteristics (Brouyère 2001, Brouyère et al. 2004a). The  
13 relative hydraulic conductivity curve shows the expected evolution with water content or  
14 suction head. For a small suction applied, fissures desaturate and the hydraulic conductivity of  
15 the chalk rock drops by several orders of magnitude. Afterwards, the water content and the  
16 relative hydraulic conductivity show a slower variation when the suction is increased. The  
17 relationship reproduces the strong reduction in hydraulic conductivity observed when the  
18 fissure network is inactive (here, a reduction by a factor 100). This validates a posteriori the  
19 use of Mualem's model together with the proposed bi-modal retention curve.

20 The way the two retention curves are combined to create the global retention curve means that  
21 until the matrix is fully saturated, fissures are empty and remain inactive. This implies that the  
22 retention model cannot accommodate by-pass flows observed when there is pressure  
23 disequilibrium between the fissures and the matrix. Therefore, the use of the model is

1 restricted to relatively slowly changing infiltration conditions. In the Hesbaye aquifer, the  
2 existence of a thick layer of loess smoothes the temporal variations of groundwater recharge  
3 rate at the top of the unsaturated chalk layer and the assumption of pressure equilibrium  
4 between the matrix and the fissures is likely to occur in the unsaturated zone (Brouyère et al.  
5 2004a). These relationships could also be used to model situations for which fast flow along  
6 fissures is observed in the unsaturated zone without pressure disequilibrium, such as water  
7 film flows (Tokunaga and Wan 1997, 1998, Tokunaga et al. 2000, Or and Tuller 2000) or  
8 fracture surface-zone flows (Tokunaga and Wan 2001). Finally, if pressure disequilibrium and  
9 by-pass flow occur, the model proposed here provides a good first approximation.

### 10 ***1.3 Partitioning the chalk porosity according to the water saturation degree***

11 Structured geological formations are often characterized by the presence of an important  
12 quantity of immobile or less mobile water located in small pores or in less pervious layers  
13 (Gerke and van Genuchten 1993). To compute contaminant transport and retardation in such  
14 formations, the dual-porosity, first-order transfer model (Coats and Smith 1964, van  
15 Genuchten and Wierenga 1976, Brouyère et al. 2000) introduces two parameters in the  
16 calibration process: the immobile water porosity  $\theta_{im}$  (dimensionless) and the first order  
17 transfer coefficient  $\alpha$  ( $T^{-1}$ ). The original form of the dual porosity model assumes that the  
18 porosity associated with the immobile water is constant. This model will be further called the  
19 “classical dual porosity model” (CDPM approach). However, when modelling transient  
20 unsaturated groundwater flows, any reduction of water content should affect the distribution  
21 of water between mobile and immobile water phases. If one of these terms is assumed  
22 constant, for example the immobile water content, when the total water content is reduced by  
23 amplitude close to that of the effective porosity, i.e. the mobile water porosity  
24  $\theta_m$  (dimensionless), the latter tends towards zero, which is physically and mathematically  
25 unacceptable (Zurmühl and Durner 1996). In reality, as discussed by Brouyère et al. (2004a),

1 under variably saturated flow conditions, when the rock mass desaturates, fissures desaturate  
2 first and the hydraulic conductivity of the rock mass is globally reduced. In such conditions,  
3 water located in the matrix cannot be considered as immobile anymore and it becomes the  
4 only vector of mobility of contaminants. In other words, matrix porosity becomes associated  
5 with the effective porosity.

6 It appears that, in order to be applicable in transient variably saturated groundwater flow  
7 conditions, the dual porosity first order transfer model has to be adjusted. Water needs to be  
8 distributed between mobile and immobile phases, depending of the degree of saturation.  
9 Zurmühlh and Durner (1996) suggested several criteria, the simplest being the consideration of  
10 a constant ratio between the immobile water porosity and the total water content:

$$11 \quad c_{part} = \frac{\theta_{im}}{\theta} \quad (7a)$$

12 They also suggested distributing the water according to the ratio of relative hydraulic  
13 conductivity values associated with the immobile water and the total water content:

$$14 \quad c_{part} = \frac{k_r(\theta_{im})}{k_r(\theta)} \quad (7b)$$

15 With this approach, the hydraulic conductivity curve reflects the distribution of velocities in  
16 the medium, analogous to a capillary bundle model (Rao et al. 1976, Toride et al. 1995). This  
17 adjusted model will be further called the “dynamic dual porosity model” (DDPM approach).  
18 Figure 4 illustrates this concept, using the hydraulic conductivity curve defined for the chalk  
19 in the previous section, with a partitioning coefficient  $c_{part} = 0.01$ . At saturation, the effective  
20 porosity of the chalk rock is small, associated with the fissures, and the quantity of immobile  
21 water in the dual porosity is high, associated with the chalk matrix (Figure 4a). When fissures  
22 desaturate, the effective porosity of the chalk is higher, associated with a part of the water  
23 moving in the matrix. At the same time, the quantity of water that is considered as immobile  
24 is reduced (Figure 4b).

## 1 **1.4 First conclusions**

2 In the variably saturated dual-porosity, dual-permeability chalk, the hydraulic conductivity is  
3 likely to change rapidly by several orders of magnitudes and the distribution of water content  
4 between mobile and immobile water needs to be continuously updated. The mathematical  
5 model presented here above meets these two essential criteria. When the fissures are partially  
6 or fully saturated, the model predicts a high hydraulic conductivity and a low effective  
7 porosity of the chalk rock. At the same time, dual-porosity effects are likely to be important  
8 because of the large porosity of the matrix. When the saturation degree is reduced, fissures  
9 become inactive and the hydraulic conductivity curve is reduced by several orders of  
10 magnitude. The effective porosity of the chalk rock becomes larger, associated with water  
11 present in the rock matrix. The next section describes the adaptations to groundwater flow and  
12 contaminant transport equations solved in the SUFT3D code (Brouyère 2001, 2003) for the  
13 integration of these concepts.

## 14 **2 Numerical modelling of groundwater flow and contaminant transport in** 15 **the variably saturated chalk**

### 16 **2.1 Variably saturated groundwater flow modelling**

17 In the finite element simulator SUFT3D, a generalized form of Richard's equation is used to  
18 model groundwater flow in variably saturated conditions:

$$19 \quad F \frac{\partial h}{\partial t} = \underline{\nabla} \cdot \left[ \left( \underline{K}_s k_r(\theta) \right) \cdot \underline{\nabla}(h+z) \right] + q \quad (8)$$

20 where  $h$  is the pressure head (L), positive in the saturated zone and negative in the  
21 unsaturated zone;  $z$  is the elevation head (L);  $\underline{K}_s$  is the saturated hydraulic conductivity  
22 tensor ( $L T^{-1}$ );  $k_r(\theta)$  is the relative hydraulic conductivity (-);  $q$  is a source/sink term ( $T^{-1}$ );  
23 and  $F$  is a generalized storage coefficient ( $L^{-1}$ ) that can be expressed as:

$$1 \quad F = S_s + \frac{d\theta}{dh} \quad (9)$$

2 In Equation (9), the first term is the specific storage coefficient  $S_s$  ( $L^{-1}$ ) that accounts for the  
 3 compressibility of water and porous medium. The second term represents the storage of water  
 4 in the unsaturated zone; it is the first derivative of the retention curve  $\theta(h)$  (see Annex 1 for  
 5 its complete mathematical expression).

6 In the saturated zone, one can generally consider that the water content is constant  
 7 ( $\theta = \theta_s = \text{constant}$ ), in which case:

$$8 \quad \frac{d\theta}{dh} = 0 \quad (10a)$$

9 In the unsaturated zone, the specific storage coefficient can often be disregarded compared to  
 10 the water storage term ( $S_s \ll d\theta/dh$ ), in which case, one can write:

$$11 \quad F \frac{\partial h}{\partial t} = \frac{d\theta}{dh} \frac{\partial h}{\partial t} = \frac{\partial \theta}{\partial t} \quad (10b)$$

12 Based on these assumptions, Equation (8) can be expressed as follows:

$$13 \quad S_s \frac{\partial h}{\partial t} + \frac{\partial \theta}{\partial t} = \nabla \cdot \left[ \underline{\underline{K_s}} k_r(\theta) \cdot \nabla (h + z) \right] + q = \nabla \cdot \underline{\underline{v_D}} + q \quad (11)$$

14 The term  $\underline{\underline{v_D}}$  in the right-hand side of Equation (11) is the Darcy flux ( $LT^{-1}$ ). The retention  
 15 curve  $\theta(h)$  and the hydraulic conductivity curve  $k_r(\theta)$  are presented in the previous sections.

16 One of the two components of the storage term (left-hand side of Equation 11) is equal to zero  
 17 depending on whether the computation is performed in the saturated zone or in the  
 18 unsaturated zone. This formulation, which distinguishes the saturated and the unsaturated  
 19 parts of the aquifer, is necessary to avoid numerical problems when applying the numerical  
 20 solution proposed by Celia et al. (1990) which, in its original form, is not suitable to simulate  
 21 a fully saturated medium.

1 For unsaturated groundwater flow, the SUFT3D code applies the formulation of Celia et al.  
2 (1990) to Equation (11) linearised using Picard, Newton-Raphson or a mixed form of these  
3 two schemes (Putti and Paniconi 1992). The numerical solution of Equation (11) is obtained  
4 by applying the control volume finite element method (e.g., Letniowski and Forsyth 1991,  
5 Therrien and Sudicky 1996). Convergence improvement is achieved using a dynamic  
6 relaxation scheme (Cooley 1983) together with the target-based full Newton time stepping  
7 scheme proposed by Diersch and Perrochet (1999). The solution to the discretized and  
8 linearized equation system is obtained using the sparse-system equation solver WatSolv  
9 (VanderKwaak et al. 1997). Details can be found in Brouyère (2001).

## 10 **2.2 Solute transport modelling**

11 In the SUFT3D code, the general mass conservation equation applied to the solute  
12 contaminant in the variably saturated chalk is:

$$13 \quad \frac{\partial(\theta_m C)}{\partial t} + \frac{\partial(\theta_{im} C_{im})}{\partial t} = -\underline{\nabla} \cdot (\underline{v}_D C) + \underline{\nabla} \cdot (\theta_m \underline{\underline{D}}_h \cdot \underline{\nabla} C) - \lambda(\theta_m C + \theta_{im} C_{im}) + qC' \quad (12)$$

14 In Equation (12), the left-hand side represents the storage of solute at a concentrations  $C$   
15 ( $\text{ML}^{-3}$ ) in the mobile water (associated porosity  $\theta_m$ ) and  $C_{im}$  ( $\text{ML}^{-3}$ ) in the immobile water  
16 and (associated porosity  $\theta_{im}$ ). In the right-hand side, the first term is the advective flux. The  
17 second term is the hydrodispersive flux ( $\underline{\underline{D}}_h$  is the hydrodynamic dispersion tensor,  $\text{L}^2\text{T}^{-1}$ ).  
18 The third term represents solute degradation in the mobile and the immobile water ( $\lambda$  is the  
19 first-order degradation constant,  $\text{T}^{-1}$ ) and the last term accounts for a source/sink, at a rate  $q$   
20 ( $\text{T}^{-1}$ ) and concentration  $C'$  ( $\text{ML}^{-3}$ ), with  $C' = C_{inj}$  if  $q > 0$  and  $C' = C$  if  $q < 0$  ( $C_{inj}$  being  
21 the concentration in the injected fluid).

1 Equation (12) is called the divergence form of the transport equation (Diersch 2001, Saaltink  
 2 et al. 2004). Expanding the mass storage and the advective flux terms in Equation (12) and  
 3 using Equation (11) gives the advective form of the transport equation:

$$\begin{aligned}
 & \theta_m \frac{\partial C}{\partial t} + \frac{\partial(\theta_{im} C_{im})}{\partial t} = \\
 & - \underline{v}_D \cdot \underline{\nabla} C + \underline{\nabla} \cdot (\theta_m \underline{D}_h \cdot \underline{\nabla} C) + q(C' - C) - \lambda(\theta_m C + \theta_{im} C_{im}) + \left( F \frac{\partial h}{\partial t} - \frac{\partial \theta_m}{\partial t} \right) C
 \end{aligned} \tag{13}$$

5 The advective formulation offers several advantages, e.g., when solving chemical reaction  
 6 problems (Huyakorn et al. 1985) or when applying lagrangian methods to solve advection-  
 7 dominated problems (Yeh 1990). It has also the advantage of facilitating the implementation  
 8 of the porosity-partitioning concept.

9 Mass conservation equation applied to the immobile water alone can be written as:

$$\frac{\partial(\theta_{im} C_{im})}{\partial t} = \alpha(C - C_{im}) - \lambda \theta_{im} C_{im} + f_{\Delta\theta}^C \tag{14}$$

11 In Equation (14),  $\alpha$  is the first-order transfer coefficient ( $T^{-1}$ ) that accounts for diffusive  
 12 solute exchange between mobile and immobile water,  $f_{\Delta\theta}^C$  represents solute exchange due to  
 13 water transfer between mobile and immobile water when the degree of saturation varies with  
 14 time ('advective' exchange).

15 Practically speaking, the computation is performed as follows. The groundwater flow  
 16 simulation is performed on a time step  $\Delta t$ . Based on the results, the variation in water content  
 17 is evaluated at each calculation point between time  $t$  and  $t + \Delta t$ . At a given calculation point,  
 18 if the water content  $\theta$  increases, the immobile water porosity  $\theta_{im}$  increases proportionally  
 19 (see Figure 4). In that case, a quantity of water containing solute at a concentration  $C$  is  
 20 "transferred" from the mobile water to the immobile water:

$$\text{If } \frac{\partial \theta_{im}}{\partial t} > 0, f_{\Delta\theta}^C = \frac{\partial \theta_{im}}{\partial t} C \approx \frac{\theta_{im}(t + \Delta t) - \theta_{im}(t)}{\Delta t} C = \frac{\Delta \theta_{im}}{\Delta t} C \tag{15a}$$

1 In contrast, if the water content  $\theta$  decreases, the immobile water porosity  $\theta_{im}$  decreases  
 2 proportionally and a quantity of water containing solute at a concentration  $C_{im}$  is transferred  
 3 from the immobile to the mobile water:

$$4 \text{ If } \frac{\partial \theta_{im}}{\partial t} < 0, f_{\Delta \theta}^C = \frac{\partial \theta_{im}}{\partial t} C_{im} \approx \frac{\theta_{im}(t + \Delta t) - \theta_{im}(t)}{\Delta t} C_{im} = \frac{\Delta \theta_{im}}{\Delta t} C_{im} \quad (15b)$$

5 Considering Equations (15a) and (15b), Equation (14) can be written in a general form:

$$6 \theta_{im} \frac{\partial C_{im}}{\partial t} = \alpha(C - C_{im}) - \lambda \theta_{im} C_{im} + (C^* - C_{im}) \frac{\Delta \theta_{im}}{\Delta t} \quad (16)$$

$$7 \text{ with } C^* = C \text{ if } \frac{\Delta \theta_{im}}{\Delta t} > 0$$

$$8 \quad C^* = C_{im} \text{ if } \frac{\Delta \theta_{im}}{\Delta t} < 0$$

9 In Equation (16), the last term of the right-hand side is only present if the transfer of water  
 10 occurs from the mobile water to the immobile water.

11 Introducing Equation (16) into Equation (13) provides the mass conservation equation applied  
 12 to the solute in the mobile water alone:

$$13 \theta_m \frac{\partial C}{\partial t} + \frac{\Delta \theta_{im}}{\Delta t} C^* = \quad (17)$$

$$- \underline{v}_D \cdot \underline{\nabla} C + \underline{\nabla} \cdot (\theta_m \underline{D}_h \cdot \underline{\nabla} C) + q(C' - C) - \lambda \theta_m C - \alpha(C - C_{im}) + \left( F \frac{\partial h}{\partial t} - \frac{\partial \theta_m}{\partial t} \right) C$$

14 It is still necessary to discuss the last term of the right-hand side of Equation (17). In the  
 15 saturated zone, the water content terms ( $\theta$ ,  $\theta_m$ ,  $\theta_{im}$ ) are assumed constant and  $F \approx S_s$ . In  
 16 that case:

$$17 \left( F \frac{\partial h}{\partial t} - \frac{\partial \theta_m}{\partial t} \right) C \approx S_s \frac{\partial h}{\partial t} C \approx S_s \frac{\Delta h}{\Delta t} C \quad (18)$$

18 Since  $S_s$  is small, this term is often negligible. In that case, Equation (17) reduces to the  
 19 “classical” solute transport equation in the presence of a dual-porosity process:



$$\theta_m \frac{\partial C}{\partial t} = -\underline{v}_D \cdot \underline{\nabla} C + \underline{\nabla} \cdot (\theta_m \underline{D}_h \cdot \underline{\nabla} C) + q(C' - C) - \lambda \theta_m C - \alpha(C - C_{im}) \quad (19)$$

2 In the unsaturated zone, one can write:

$$\theta = \theta_m + \theta_{im} + \theta_c \quad (20)$$

4 where  $\theta_c$  represents a possible portion of isolated water, which does not contribute at all to  
5 transport processes. This quantity of water can be assumed as either negligible or invariant  
6 with time, for which case one can write,

$$F \frac{\partial h}{\partial t} \approx \frac{\partial \theta}{\partial t} = \frac{\partial \theta_m}{\partial t} + \frac{\partial \theta_{im}}{\partial t} \approx \frac{\Delta \theta_m}{\Delta t} + \frac{\Delta \theta_{im}}{\Delta t} \quad (21)$$

8 Introducing Equation (21) into Equation (17) provides the mathematical form of the transport  
9 equation in the mobile water, in the unsaturated zone:

$$\theta_m \frac{\partial C}{\partial t} = -\underline{v}_D \cdot \underline{\nabla} C + \underline{\nabla} \cdot (\theta_m \underline{D}_h \cdot \underline{\nabla} C) + q(C' - C) - \lambda \theta_m C - \alpha(C - C_{im}) + \frac{\Delta \theta_{im}}{\Delta t} (C - C^*) \quad (22)$$

11 The last term of the right-hand side exists when the transfer of water occurs from the  
12 immobile water to the mobile water (i.e.  $\Delta \theta_{im} / \Delta t < 0$ ,  $C^* = C_{im}$ ), the solute present in the  
13 mobile water being diluted or concentrated depending whether  $C > C_{im}$  or not. If water is  
14 transferred from the mobile to the immobile water, the concentration in the mobile water is  
15 not affected by loss of a portion of this phase.

16 Different numerical approaches can be found in the literature for the solution of the dual-  
17 porosity, first-order transfer model, differing from a point of view of computational efficiency  
18 and stability (Gallo et al. 1996). A first solution is to use a fully coupled approach (Gambolati  
19 et al. 1994) for which the concentration in the immobile water  $C_{im}^*$  is considered as a state  
20 variable just like the concentration in the mobile water  $C$ . The number of unknowns changes  
21 from  $N$  to  $2N$  ( $N$  being the number of nodes in the discretization), which increases  
22 considerably the size of the equation system to be solved and therefore the demand on

1 memory and computation time. Decoupled approaches give very accurate results at lower  
2 computational costs. Gambolati et al. (1993) use an integro-differential approach that allows  
3 an analytical computation of concentrations in the immobile water over a time step. The  
4 resulting expression is then back-substituted into the equation relative to the concentration  
5 evolution in the mobile water. Unfortunately, this numerical scheme shows instabilities if the  
6 first order transfer coefficient between mobile and immobile water becomes large. Ibaraki and  
7 Sudicky (1995) or Gambolati et al. (1996) propose a decoupled approach for which the  
8 differential equation relative to mass-conservation in the immobile water (Equation 14) is  
9 approximated by a finite difference scheme written over the computation time step. Once  
10 more, the resulting expression is back-substituted into the equation describing the evolution of  
11 concentration in the mobile water (Equation 17). This approach gives very accurate results  
12 and shows a very good stability.

13 A fourth approach proposed by Biver (1993) is used in the SUFT3D code. A semi-analytical  
14 expression is found for the concentration evolution in the immobile water (Equation 16) over  
15 the computation time step. The resulting expression is substituted in Equation 17. The  
16 resulting partial differential equation is then solved using the finite element method. Details  
17 can be found in Brouyère (2001). The coding of the dual-porosity concept in the SUFT3D  
18 code was verified by comparison with the analytical solutions proposed by van Genuchten  
19 and Wierenga (1976), using the CXTFIT code (Toride et al. 1995) and by comparison with  
20 FRAC3DVS (Therrien 1992, Therrien and Sudicky 1996) based on the computation of a  
21 synthetic radially converging tracer experiment (Brouyère et al. 2000).

### 22 **3 Modelling tracer experiments performed in the chalk**

23 Details about the tracer experiments performed in the unsaturated zone overlying the Hesbaye  
24 aquifer in Belgium can be found in Brouyère (2001) and Brouyère et al. (2004a). After a short

1 description of the experimental setup, the simulations of the experiments performed with the  
2 SUFT3D code are described and discussed.

### 3 **3.1 *Tracer experiments performed in the chalk***

4 In the Hesbaye region, the geological succession consists, from top to bottom, of 13 m of  
5 loess formations, 2 to 4 m of flint conglomerate and 32 m of fissured chalk. The aquifer,  
6 located in the fissured chalk, is unconfined. The aquifer basis is formed by several meters of  
7 smectite clay (Brouyère et al. 2004b). The experimental site, located at Bovenistier, is  
8 equipped with 7 boreholes drilled and screened at different depths in the saturated zone (the  
9 central well PC and two piezometers Pz CS and Pz 12) and in the unsaturated zone (Pz CNS  
10 in the unsaturated chalk, Pz CGL in the flint conglomerate, Pz LB and Pz LS in the loess).  
11 Undisturbed core samples were collected during the drilling of the boreholes for laboratory  
12 measurements, such as hydraulic conductivity measurements, the determination of  
13 unsaturated properties of the different geological formations and the analysis of nitrate and  
14 pesticide contents. In the field, infiltration tests were performed in the unsaturated zone and  
15 pumping tests were performed in the saturated zone. Tracer experiments were performed in  
16 both the unsaturated and saturated zone. Table 1 summarizes the information relative to these  
17 tracer injections.

18 Two tracer tests were performed between Pz CS and the central well in the saturated zone of  
19 the chalk aquifer, in radially converging flow conditions. For the first injection (phase 1 in  
20 Table 1), eosin Y was used and the pumping rate at the recovery well was 1.2 m<sup>3</sup>/h. For the  
21 second injection (phase 2 in Table 1), naphthionate was used and the pumping rate at the  
22 recovery well was 6 m<sup>3</sup>/h. Figure 5 presents the measured breakthrough curves at the central  
23 well PC.

24 Two tracer injections were performed in the unsaturated chalk, from Pz CNS (Figure 6). For  
25 the first injection with potassium chloride (KCl), artificial recharge conditions were created

1 by adding water at a constant rate ( $0.3 \text{ m}^3/\text{h}$ ) in the well after tracer injection. This led to  
2 enhanced hydraulic gradient between the injection point and the aquifer and locally to a  
3 higher degree of saturation in the unsaturated chalk. As the chalk layer is overlain by a thick  
4 loess formation, actual recharge conditions are likely to be less intensive. A second tracer  
5 injection was performed using potassium iodide (KI), without addition of water after tracer  
6 injection. This configuration reflects better the actual seepage conditions in the unsaturated  
7 chalk. The first tracer experiment in the unsaturated chalk was performed during a period of  
8 low groundwater levels that prevailed during the whole tracer monitoring period (3 days). For  
9 this experiment, the unsaturated thickness of chalk crossed by the tracer was approximately  
10 8.5 m. For the second tracer experiment in the unsaturated chalk, the iodide tracer was also  
11 injected at the period of low groundwater levels. However, the duration of the test was longer  
12 (about two years) and a progressive rise in groundwater levels was observed in the aquifer  
13 during that period. The unsaturated thickness of chalk, so the distance travelled by the tracer,  
14 was reduced down to 4 to 3 m.

### 15 **3.2 Conceptual model and discretisation**

16 Three finite element meshes were used for modelling exercise: a regional mesh and two local  
17 meshes, both centred around the central well and refined in the area of the tracer experiments.  
18 The regional model was developed using a finite element mesh with an horizontal extension  
19 of  $320 \text{ m} \times 320 \text{ m}$  and a vertical dimension of 32 m representing the saturated zone of the  
20 aquifer. This model, used for calculating groundwater heads at the boundaries of the two local  
21 meshes, was calibrated for natural groundwater flow conditions, based on water levels  
22 measured in the region during the period of the experiments and on pumping tests performed  
23 in the central well (results not shown). In order to account for the rock alteration and strain  
24 relaxation at shallow depth, values of hydraulic conductivity defined in the upper layers are  
25 higher (between  $1.0 \times 10^{-4}$  and  $4.0 \times 10^{-4} \text{ m.s}^{-1}$ ) than in the deepest layers (between  $5.0 \times 10^{-5}$  and

1  $1.0 \times 10^{-4} \text{ m.s}^{-1}$ ). The hydraulic conductivity field obtained from the calibration of the regional  
2 model was transferred to the corresponding finite elements of the local meshes.

3 Tracer tests performed in the saturated zone were modelled using a finite element mesh with  
4 an horizontal extension of  $80 \text{ m} \times 80 \text{ m}$  and a vertical dimension of  $32 \text{ m}$  representing the  
5 saturated zone of the chalk aquifer (10 layers of finite elements ranging from  $2 \text{ m}$  to  $5 \text{ m}$ ).

6 Tracer tests performed in the unsaturated chalk were modelled using another mesh with a  
7 horizontal extension of  $18 \text{ m} \times 18 \text{ m}$ , extending over the whole thickness of the saturated and  
8 unsaturated zone ( $50 \text{ m}$ ). In the unsaturated zone, the discretisation is made of  $1 \text{ m} \times 1 \text{ m} \times 1$   
9  $\text{m}$  cubic finite elements.

10 At the lateral boundaries of the local meshes, prescribed heads (Dirichlet boundary  
11 conditions) were defined in the saturated zone. Head values were set equal to those computed,  
12 at the same location, using a “regional” model run with the same stress factors (recharge,  
13 pumping rate...). In the unsaturated zone, water fluxes being essentially vertical, lateral faces  
14 were considered as impervious. At the top of the model, a flux-type (Neumann) boundary  
15 condition was prescribed. The transient nature of recharge conditions was not considered and  
16 a mean annual rate of  $9.5 \times 10^{-9} \text{ m/s}$  was applied (approx.  $300 \text{ mm/year}$ ).

17 The characteristic curves  $\theta(h)$  and  $k_r(\theta)$  were defined for each geological formation  
18 encountered, together with an estimation of the saturated hydraulic conductivity  $K_s$ . For the  
19 loess formations, van Genuchten relationships were used, considering the mean estimates of  
20 the parameters obtained by fitting this relationship on retention curves that were measured on  
21 loess samples (Brouyère et al. 2004a). In the absence of data, the flint conglomerate was  
22 considered as a silty sand for which mean van Genuchten parameters were taken from tables  
23 provided by Carsel and Parrish (1988). For the chalk, a bimodal relationship similar to that  
24 established in section 2.1 was considered. Parameter values considered for the different  
25 geological formations are summarized in Table 2.

1 Saturated hydraulic conductivity values for the loess and conglomerate formations were  
2 obtained by modelling constant head infiltration tests performed in the field. Calibrated  
3 hydraulic conductivity values are in good agreement with those derived from infiltration tests  
4 (Brouyère 2001, Brouyère et al. 2004a):  $1.5$  to  $5.0 \times 10^{-7}$  m/s for the loess formation and  
5  $2.0 \times 10^{-6}$  m/s for the flint conglomerate. For the unsaturated chalk, the infiltration tests  
6 provided an estimation of saturated hydraulic conductivity that was considered too low  
7 compared to values derived from the pumping tests. It was decided to use a mean hydraulic  
8 conductivity  $K_s = 1.0 \times 10^{-4}$  m/s.

### 9 **3.3 Modelling tracer experiments performed in the saturated chalk**

10 The two tracer experiments performed between Pz CS and the central well were modelled  
11 with the SUFT3D code, using the CDPM approach. Figure 5 shows the comparison between  
12 measured and fitted breakthrough curves. Table 3 summarizes adjusted hydrodispersive  
13 parameters.

14 The calibrated effective porosity is low, typical for the fissure porosity of the chalk. First  
15 arrivals are not perfectly calibrated because the mesh is not refined enough for modelling the  
16 low dispersive transport along fissures between Pz CS and the central well. Because of  
17 numerical dispersion, reducing further the value of longitudinal dispersivity  $\alpha_L$  (L) did not  
18 have any effect on the computed breakthrough curve. The calibrated longitudinal dispersivity  
19 coefficient is thus probably overestimated. The calibrated first-order transfer coefficient  $\alpha$  is  
20 in a good agreement with values found for other tracer experiments performed in the fissured  
21 chalk in the Hesbaye aquifer (Biver 1993, Hallet 1999). The porosity associated with the  
22 immobile water  $\theta_{im}$  is relatively small. However, in the saturated zone, the dual-porosity  
23 concept is usually considered as a diffusive process characterised by relatively slow kinetics.  
24 Because the tracers were injected very close to the recovery well, the mean travel time in the

1 aquifer is very short. It is thus likely that a part of the immobile water does not take part in the  
2 retardation process. The calibrated immobile water porosity thus reflects the quantity of water  
3 that is involved in the physical retardation process rather than the total quantity of immobile  
4 water.

### 5 ***3.4 Modelling tracer experiment performed in the unsaturated chalk under artificial*** 6 ***injection conditions***

7 Steady state conditions were assumed for modelling the tracer experiment in the unsaturated  
8 chalk for high recharge rates. Tracer injection and water recharge conditions were modelled  
9 considering a nodal source term applied at the finite element node which is the nearest from  
10 the screen level of Pz CNS in the three-dimensional finite element mesh. A constant rate of  
11  $0.3 \text{ m}^3/\text{h}$  was applied in the model. A tracer injection duration of one hour was considered  
12 (close to the actual 53 min. that were actually needed to inject the 300 litres of tracer fluid).  
13 The calibration was based on the fitting of the chloride breakthrough curve, this tracer being  
14 considered as conservative (i.e. no sorption or degradation). First, the CDPM approach was  
15 considered, assuming that effective porosity and immobile water porosity are constant values  
16 defined at the R.E.V. scale, independent of the rock saturation. In this case, the degree of  
17 saturation plays only a role for the computation of Darcy fluxes. The fitting parameters are the  
18 effective porosity, the longitudinal dispersivity, the immobile water porosity and the first  
19 order transfer coefficient between mobile and immobile water phases. Second, the DDPM  
20 approach, as described in the previous sections, was considered. In this case, the saturation  
21 degree of the rock plays a role on both the computation of Darcy fluxes (groundwater flow  
22 computation) and on the distribution of water between mobile and immobile porosity  
23 (transport computation). The magnitudes of effective porosity and immobile water porosity  
24 are not exactly known. The only information is the partitioning coefficient  $c_{part}$  considered as  
25 a fitting parameter together with the longitudinal dispersivity and the first-order transfer

1 coefficient. Figure 6a presents the chloride breakthrough curve measured in the field together  
2 with the best fit obtained (by trial-and-error) considering the two approaches. Table 3  
3 summarizes the calibrated parameters.

4 For the CDPM approach, fitted parameters are close to those obtained from modelling the  
5 tracer experiments performed in the saturated zone (see Table 3). The effective porosity is low  
6 (fissure porosity) and the longitudinal dispersivity coefficient is relatively small but probably  
7 overestimated again. The immobile water porosity is similar to that considered in the  
8 saturated zone, reflecting some bypass of the immobile water during the tracer downward  
9 migration in the unsaturated chalk. The first order transfer coefficient is very similar to that  
10 obtained in the saturated zone. This indicates that hydrodispersive processes governing the  
11 mobility of tracers during these experiments were similar to those prevailing in the saturated  
12 zone: a fast, preferential migration of tracers along partially or fully saturated fissures and a  
13 strong physical retardation in the immobile water located in the matrix porosity.

14 When the DDPM approach is considered, a partitioning coefficient  $c_{part} = 0.02$  is found. This  
15 indicates that the effective porosity is small and the immobile water porosity is large (see  
16 Figure 4). At the same time, the calibrated first order transfer coefficient is smaller than that  
17 obtained from the CDPM approach. In fact, the degree of saturation varies spatially and  
18 therefore also the distribution of porosity between mobile and immobile water and the  
19 hydraulic conductivity. For lower degrees of saturation, the effective porosity is larger. This  
20 contributes to the retardation of that portion of the tracer migrating across these zones, as  
21 compared to the quantity of tracer migrating across zones characterised by a higher degree of  
22 saturation. The DDPM approach thus requires less ‘diffusive’ effect for modelling tracer late  
23 arrivals compared to the CDPM approach. At the extreme, delayed tracer arrivals could be  
24 obtained just by adapting the hydraulic conductivity functionality and the partitioning  
25 coefficient until the breakthrough curve is reproduced independently from any dual-porosity



1 effect. However, it is not realistic to think that diffusion into the immobile water does not  
2 have any influence on the mobility and retardation of tracers in the chalk.

3 Whatever the modelling approach that is considered (CDPM or DDPM), modelling results  
4 confirm the conceptual model of tracer transport for intensive recharge conditions. The fissure  
5 network was active, driving tracers at high velocities across the unsaturated zone, together  
6 with an important dual-porosity effect that produced physical retardation due to storage of  
7 tracers in the immobile water located in the low permeability matrix. It has to be mentioned  
8 that previous attempts, consisting of modelling the unsaturated properties of the chalk with  
9 ‘classical’ functionalities (such as unimodal van Genuchten relationships) were unable to  
10 explain and reproduce short travel times such as those observed in the field (Crispin 2000).

11 The bimodal functionality developed to model the behaviour of the variably saturated chalk is  
12 thus one of the key factors that allow modelling of the tracer experiments. It allows a fair  
13 reproduction of the very dynamic behaviour of the variably saturated chalk when the  
14 saturation degree varies. A full dual-permeability approach, modelling the chalk by two flow  
15 continua (one for the matrix, one for the fissured component) like that of Gerke and van  
16 Genuchten (1993) could probably also provide an even better representation of the  
17 phenomena but the required data are generally not available.

### 18 ***3.5 Modelling of the tracer experiment performed in the unsaturated chalk for natural*** 19 ***infiltration conditions***

20 This tracer experiment is more representative of natural downward migration conditions of  
21 solute contaminants in the unsaturated chalk. Because the tracer experiment lasted for two  
22 years, the interpretation and modelling should ideally consider the transient nature of recharge  
23 conditions and the variations in piezometric levels. However, for computation time and  
24 numerical reasons, simplified assumptions were considered. First, as no detailed data were  
25 available on the evolution of water contents with time in the unsaturated zone, the transient

1 nature of the recharge was neglected. Second, as mentioned previously, in order to save  
2 memory and computation time, a local model was needed for the transport simulations. At the  
3 limits of this local model, boundary conditions were not easy to define in the fluctuating zone  
4 of the groundwater table. Normally, boundary conditions should be switched from impervious  
5 to prescribed head as the groundwater table rises, but this was not available in the SUFT3D.  
6 In a first attempt, the rise of the groundwater table was considered by increasing the total head  
7 prescribed at the Dirichlet nodes. With such a configuration, water could enter or leave the  
8 model via the bottom layers only, resulting in very artificial Darcy fluxes and pressure fields,  
9 with the direct consequence that numerical problems were inevitable in the transport  
10 simulations.

11 Steady state groundwater flow simulations were finally performed, considering groundwater  
12 levels prevailing at the beginning of the tracer experiments (low groundwater levels). For the  
13 transport simulations, the size of the time steps was restricted to 1 hour because of the  
14 numerical Courant criterion. With the total simulation time being 2 years, very long  
15 computation times of approximately one week were required.

16 Because of the above restrictions, the objectives of the transport simulations performed in  
17 natural infiltration conditions were to check: (1) if the strong contrast observed in terms of  
18 travel times across the unsaturated chalk could be reproduced; (2) to check the consistency of  
19 the conceptual model postulating that contaminants migrate across the matrix in natural  
20 recharge conditions; and (3) to have an estimation of the effective porosity associated to the  
21 chalk in natural infiltration conditions. For the simulations, unless stated here after,  
22 hydrodispersive parameters were set equal to those obtained when modelling the tracer  
23 experiments performed under intense injection conditions.

1 Using the DDPM approach with the hydrodispersive parameters obtained for the tracer  
2 experiments performed under experimental intense injection conditions did not result in any  
3 significant tracer arrival during the sampling campaign (almost 2 years).

4 Different simulations were performed considering the CDPM approach. It was first  
5 considered that all the water located in the chalk was mobile (no dual-porosity effect).  
6 Practically, a value of 35% was defined for the effective porosity. As for the DDPM  
7 approach, the result of this simulation did not show any significant tracer arrival at the  
8 pumping well during the period corresponding to the sampling campaign. The second test  
9 performed with the CDPM approach consisted in trying to obtain a modelled first arrival of  
10 the order of the observed one, by adjusting the effective and the immobile water porosity.  
11 This was obtained by using an effective porosity of 10% and an immobile water porosity of  
12 15% in the model (Figure 6b). The order of magnitude of the first arrival is respected (almost  
13 one year). However, the modelled breakthrough curve shows an evolution that departs from  
14 the measured one, the latter being sharper. This is probably related to the fact that tracer  
15 arrivals were not only explained by the downward migration of the tracer in the unsaturated  
16 zone, but also by the rise of the groundwater table that washed the tracer during its downward  
17 migration in the unsaturated zone.

18 The effective porosity (10%) obtained by adjusting the first arrival is too small to be  
19 representative of the matrix porosity. However, as stated before, groundwater levels  
20 prevailing at the beginning of the tracer experiment were considered in the simulations.  
21 Because of that, the travel distance in the model (approximately 9m) is almost 3 times the  
22 actual distance covered by the tracer (approximately 3m). A simple rule of three thus provides  
23 a rough estimation of 30% for the effective porosity from the 10% obtained with the model,  
24 which is in better agreement with the estimated chalk matrix porosity.

1 From the analysis of these simulations, it can be concluded that, for natural infiltration  
2 conditions, solute contaminant downward migration occurs across the chalk matrix, as  
3 suggested in the conceptual model proposed by Brouyère et al. (2004a).

#### 4 **4 General conclusions**

5 This research has provided several very useful contributions. First, a new functionality is  
6 proposed for modelling the fundamental unsaturated properties (retention curve and hydraulic  
7 conductivity curve) for structured formations such as fissured rocks or macroporous soils.  
8 This functionality allows an analytical evaluation of the hydraulic conductivity curve from the  
9 retention curve and it is continuously derivable over the whole range of water contents. It  
10 reproduces naturally the fast and strong variation in hydraulic conductivity that is often  
11 observed in structured media when the preferential flow paths become desaturated. Second, a  
12 generalisation of the dual-porosity model is proposed to the case of variably saturated  
13 groundwater flows, in a way that is similar to that proposed by Zurmühl and Durner (1996).  
14 Third, the mathematical formulation and the numerical implementation of all these  
15 developments in the three-dimensional groundwater flow and transport simulator SUFT3D  
16 are described. All these developments are used to provide an innovative and unified point of  
17 view on the hydrodynamic and hydrodispersive behaviour of variably saturated dual-porosity,  
18 dual-permeability chalk.

19 The modelling results obtained for the different tracer experiments performed in the saturated  
20 and the unsaturated chalk at Bovenistier under various experimental conditions confirm that  
21 the conceptual model proposed by Brouyère et al. (2004a) is valid to explain the  
22 hydrodispersive behaviour of the chalk for variably saturated conditions. In the saturated zone  
23 and in the unsaturated zone when intensive recharge conditions prevail, the order of  
24 magnitude of the hydrodispersive parameters shows that the fissure network drives the  
25 mobility of contaminants. At the same time, the matrix acts as a buffer where these

1 contaminants are temporarily stored and retarded. Under natural infiltration conditions, the  
2 fissure network is inactive and tracers migrate slowly downward across the high porosity, low  
3 permeability chalk matrix.

4 The analysis and interpretation of the tracer experiments performed in the chalk could be  
5 improved by considering several factors that were neglected, such as tracer density effects, the  
6 transient nature of the groundwater recharge and variations in groundwater levels. It could be  
7 interesting to perform further numerical simulations using the DDPM approach for explaining  
8 the tracer experiments performed in natural recharge conditions. A comparison between the  
9 present approach and multi-continuum approaches could provide a better understanding of the  
10 possible influence of pressure disequilibrium between the fissures and the matrix (bypass  
11 flow) on the mobility and retardation of contaminants.

12 In the field, a better understanding and quantification of the behaviour of the chalk under  
13 variably saturated flow could be obtained by performing tracer injections under variable  
14 recharge conditions. In order to monitor the impact of the fissure network inactivation, it  
15 could be interesting to perform tracer injections with variable artificial recharge rates or to  
16 perform a long duration tracer injection during which the artificial recharge is stopped at a  
17 time when tracer arrivals are still clearly above the detection limit. This kind of experiment  
18 would be similar to flow interruption methods such as proposed by Brusseau et al. (1989) or  
19 Cote et al. (2000), used to quantify kinetic retardation processes affecting the behaviour of  
20 solutes in column experiments. Such experiments could contribute to a better assessment of  
21 the functionality used for porosity partitioning and to a better understanding of fissure-matrix  
22 interactions.

23 The model presented here could be used to study the impact of groundwater table seasonal  
24 variations on the evolution of contaminant concentrations such as nitrates. This could be of  
25 direct interest for the estimation of groundwater quality trends (seasonal detrending). The

1 model is also compatible with the research works of Tokunaga and Wan (1997), Tokunaga et  
2 al. (2000), Price et al. (2000) or Tokunaga and Wan (2001). The developments presented here  
3 could thus be adapted to provide the mathematical framework for these researches.

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#### 10 **References**

- 11 Bai, M., Elsworth, D., Roegiers, J.-C., Multiporosity/multipermeability approach to the  
12 simulation of naturally fractured reservoirs, *Water Resour. Res.*, 29(6), 1621-1633, 1993.
- 13 Berkowitz, B., Bear, J., Braester, C., Continuum models for contaminant transport in  
14 fractured porous formations, *Water Resour. Res.*, 24(8), 1225-1236, 1988.
- 15 Berkowitz, B., Characterizing flow and transport in fractured geological media: A review,  
16 *Adv. Water Resour.*, 25, 861-884, 2002.
- 17 Biver, P., Etude phénoménologique et numérique de la propagation de polluants miscibles  
18 dans un milieu à porosité multiple (phenomenological and numerical study of miscible  
19 pollutant transport in a multi-porosity medium), in French, PhD thesis, University of  
20 Liège, Faculty of Applied Sciences, 389p, 1993.
- 21 Brouyère, S., Dassargues, A., Therrien, R., Sudicky, E.A., Modelling of dual porosity media :  
22 comparison of different techniques and evaluation on the impact on plume transport  
23 simulations, IAHS Publication nb 265, Ed. F.Stauffer, W.Kinzelbach, K.Kovar &  
24 E.Hoehn. IAHS Press, Wallingford, Oxfordshire, UK, 22-27, 2000.

1 Brouyère, S., Etude et modélisation du transport et du piégeage des solutés en milieu  
2 souterrain variablement saturé (study and modelling of transport and retardation of solutes  
3 in variably saturated media), in French, PhD Thesis, University of Liège, Faculty of  
4 Applied Sciences, 572p, 2001.

5 Brouyère, S., Modeling tracer injection and well-aquifer interactions: a new mathematical and  
6 numerical approach, *Water Resour. Res.*, 39(3), doi:10.1029/2002WR001813, 2003.

7 Brouyère, S., Dassargues, A., Hallet, V., Migration of contaminants through the unsaturated  
8 zone overlying the Hesbaye chalky aquifer in Belgium : a field investigation, *J. Contam.*  
9 *Hydrol.*, vol. 72 (1-4), 135-164, doi: 10.1016/j.conhyd.2003.10.009, 2004a.

10 Brouyère, S., Carabin, G., Dassargues, A., Climate change impacts on groundwater resources:  
11 modelled deficits in a chalky aquifer, Geer basin, Belgium, *Hydrogeol. J.*, 12, 123-134,  
12 doi: 10.1007/s10040-003-0293-1, 2004b.

13 Brusseau, M. L., Rao, P. S. C., Jessup, R. E., Davidson, J. M., Flow interruption: a methods  
14 for investigating sorption nonequilibrium, *J. Contam. Hydrol.*, 4, 223-240, 1989.

15 Carsel, R. F., Parrish, R. S., Developing joint probability distributions of soil water retention  
16 characteristics, *Water Resour. Res.*, 24(5), 755-769, 1988.

17 Celia, M. A., Bouloutas, E. T., Zarba, R. L., A general mass-conservative numerical solution  
18 for the unsaturated flow equation, *Water Resour. Res.*, 26(7), 1483-1496, 1990.

19 Coats, K. H., Smith, B. D., Dead-end pore volume and dispersion in porous media, *Soc. Pet.*  
20 *Eng. J.*, 4, 73-84, 1964.

21 Cooley, R. L., Some procedures for numerical solution of variably saturated flow problems, ,  
22 *Water Resour. Res.*, 19(5), 1271-1285, 1983.

23 Cote, C. M., Bristow, K. L., Ross, P. J., Increasing the efficiency of solute leaching : impacts  
24 of flow interruption with drainage of the “preferential flow paths”, *J. Contam. Hydrol.*, 43,  
25 191-209, 2000.

- 1 Crispin, C., Modélisation des essais de traçage en Hesbaye : transport des polluants en zone  
2 saturée et non saturée (Modelling tracer experiments in the Hesbaye aquifer : pollutant  
3 transport in the saturated and unsaturated zones), in French, Master thesis (D.E.S.),  
4 University of Liège, Faculty of Applied Sciences, 84p, 1999.
- 5 Diersch, H.-J.G., Perrochet, P., On the primary variable switching technique for simulating  
6 unsaturated-saturated flows, *Adv. Water Resour.*, 23, 271-301, 1999.
- 7 Diersch, H.J.G., About the difference between the convective form and the divergence form  
8 of the transport equation. In: *FEFLOW White Papers, Vol. I*, Berlin: WASY Ltd, 119-130  
9 [Chapter 6], 2001.
- 10 Durner, W., Hydraulic conductivity estimation for soils with heterogeneous pore structure,  
11 *Water Resour. Res.*, 30(2), 211-223, 1994.
- 12 Dykhuizen, R. C., Transport of solutes through unsaturated fractured media, *Water Res.*,  
13 21(12), 1531-1539, 1987.
- 14 Dykhuizen, R. C., A new coupling term for dual-porosity models, *Water Resour. Res.*, 26(2),  
15 351-356, 1990.
- 16 Finsterle, S., Using continuum approach to model unsaturated flow in fractured rock, *Water*  
17 *Resour. Res.*, 36(8), 2055-2066, 2000.
- 18 Fredlung, D. G., Xing, A., Equations for the soil-water characteristic curve, *Can. Geotech. J.*,  
19 31, 521-532, 1994.
- 20 Fredlung, D. G., Xing, A., Huang, S., Predicting the permeability function from unsaturated  
21 soils using the soil-water characteristic curve, *Can. Geotech. J.*, 31, 533-546, 1994.
- 22 Gallo, C., Paniconi, C., Gambolati, G., Comparison of solution approaches for the two-  
23 domain model of nonequilibrium transport in porous media, *Adv. Water Resour.*, 19(4),  
24 241-253, 1996.



- 1 Gambolati, G., Putti, M., Paniconi, C., Mass transfer analysis in sorbing porous media by an  
2 integro-differential approach, *Advances in Hydro-Science and –Engineering*, 1, 1819-  
3 1828, 1993.
- 4 Gambolati, G., Gallo, C., Paniconi, C., Numerical integration methods for the dual porosity  
5 model in sorbing porous media, *Comput. Methods in Water Resour.*, X, ed. A. Peters et  
6 al., Kluwer Academic Press, Dordrecht, NL, 621-628, 1994.
- 7 Gambolati, G., Putti, M., Paniconi, C., Projection methods for the finite element solution of  
8 the dual-porosity model in variably saturated porous media, in: *Advances in Groundwater  
9 Pollution Control and Remediation*, NATO-ASI Series 2: Environment, vol.9, ed. M. M.  
10 Aral, Kluwer Academic, Dordrecht, Holland, 97-125, 1996.
- 11 Gerke, H. H., van Genuchten, M. TH., A dual-porosity model for simulating the preferential  
12 movement of water and solutes in structured porous media, *Water Resour. Res.*, 29(2),  
13 305-319, 1993.
- 14 Gwo, J.-P., Multi-region flow and transport in subsurface media, Ph.D. thesis, Dpt of Civil  
15 and Environmental Engineering, Pennsylvania State University, 283p, 1992.
- 16 Hallet, V., Etude de la contamination de la nappe aquifère de Hesbaye par les nitrates:  
17 hydrogéologie, hydrochimie et modélisation mathématique des écoulements et du  
18 transport en milieu saturé (Contamination of the Hesbaye aquifer by nitrates:  
19 hydrogeology, hydrochemistry and mathematical modeling), in French, PhD Thesis,  
20 University of Liège, Faculty of Sciences, 361p, 1999.
- 21 Huyakorn, P. S., Mercer, J. W., Ward, D. S., Finite element matrix and mass balance  
22 computational schemes for transport in variably saturated porous media, *Water Resour.  
23 Res.*, 21(3), 346-358, 1985.

1 Ibaraki, M., Sudicky, E. A., Colloid-facilitated contaminant transport in discretely fractured  
2 porous media. I: Numerical formulation and sensitivity analysis, *Water Resour. Res.*,  
3 31(12), 2945-2960, 1995.

4 Letniowski, F. W., Forsyth, P. A., A control volume finite element method for three-  
5 dimensional NAPL groundwater contamination, *Int. J. Num. Met. Fl.*, 13, 955-970, 1991.

6 Mallants, D., Tseng, P.-H., Toride, N., Timmerman, A., Feyen, J., Evaluation of multimodal  
7 hydraulic functions in characterizing a heterogeneous field soil, *J. Hydrol.*, 195, 172-199,  
8 1997.

9 Mualem, Y., A new model for predicting the hydraulic conductivity of unsaturated porous  
10 media, *Water Resour. Res.*, 12(3), 513-522, 1976.

11 Or, D., Tuller, M., Flow in unsaturated fractured porous media: hydraulic conductivity of  
12 rough surfaces, *Water Resour. Res.*, 36(5), 1165-1177, 2000.

13 Othmer, H., Diekkrüger, B., Kutilek, M., Bimodal porosity unsaturated hydraulic  
14 conductivity, *Soil Sci.*, 152(3), 139-150, 1991.

15 Paniconi, C., Aldama, A. A., Wood, E. F., Numerical evaluation of iterative and noniterative  
16 methods for the solution of the nonlinear Richards equation, *Water Resour. Res.*, 27(6),  
17 1147-1163, 1991.

18 Paniconi, C., Putti, M., A comparison of Picard and Newton iteration in the numerical  
19 solution of multidimensional variably saturated flow problems, *Water Resour. Res.*,  
20 30(12), 3357-3374, 1994.

21 Peters, R. R., Klavetter, E. A., A continuum model for water movement in an unsaturated  
22 fractured rock mass, *Water Resour. Res.*, 24(3), 416-430, 1988.

23 Price, M., Downing, R. A., Edmunds, W. M., The chalk as an aquifer, in: *The hydrogeology*  
24 *of the chalk of north-west Europe*, Ed. Downing, Price, Jones, Oxford Science Publ.,  
25 Clarendon Press, Oxford, U.-K., 35-58, 1993.

- 1 Price, M., Low, R. G., McCann, C., Mechanisms of water storage and flow in the unsaturated  
2 zone of the Chalk aquifer, *J. Hydrol.*, 233, 54-71, 2000.
- 3 Pruess, K., Faybishenko, B., Bodvarsson, G.S., Alternative concepts and approaches for  
4 modelling flow and transport in thick unsaturated zones of fractured rocks, *J. Contam.*  
5 *Hydrol.*, 38, 281-322, 1999.
- 6 Putti, M., Paniconi, C., Evaluation of Picard and Newton iteration schemes for three-  
7 dimensional unsaturated flow, in *Proceedings of the IX International Conference on*  
8 *Computational Methods in Water Resources*, vol. 1, Numerical Methods in Water  
9 Resources, edited by G. Gambolati, A. Rinaldo, C. A. Brebbia, W. G. Gray, and G. F.  
10 Pinder, pp. 529-536, Computational Mechanics Publications, Billerica, Mass., 1992.
- 11 Rao, P. S. C., Green, R. E., Ahuja, L. R., Davidson, J. M., Evaluation of a capillary bundle  
12 model for describing solute dispersion in aggregated soils, *Soil Sci. Soc. Am. J.*, 40, 815-  
13 820, 1976.
- 14 Ross, P. J., Smettem, R. J., Describing soil hydraulic properties with sums of simple  
15 functions, *Soil Sci. Soc. Am. J.*, 57, 26-29, 1993.
- 16 Saaltink, M. W., Carrera, J., Olivella S., Mass balance errors when solving the convective  
17 form of the transport equation in transient flow problems, *Water Resour. Res.*, 40,  
18 W05107, doi:10.1029/2003WR002866, 2004.
- 19 Smettem, K. R. J., Kirkby, C., Measuring the hydraulic properties of a stable aggregated soil,  
20 *J. Hydrol.*, 117, 1-13, 1990.
- 21 Therrien, R., Three-dimensional analysis of variably-saturated flow and solute transport in  
22 discretely-fractured porous media, Ph.D. thesis, University of Waterloo, Dpt of Earth  
23 Sciences, 1992.
- 24 Therrien, R., Sudicky, E. A., Three-dimensional analysis of variably-saturated flow and solute  
25 transport in discretely-fractured porous media, *J. Contam. Hydrol.*, 23, 1-44, 1996.

1 Tokunaga, T. K., Wan, J., Water film flow along fracture surfaces of porous rock, *Water*  
2 *Resour. Res.*, 33(6), 1287-1295, 1997.

3 Tokunaga, T. K., Wan, J., Mechanisms for fast flow in unsaturated fractured rock, in: 5th  
4 International Symposium on reservoir wettability and its effect on oil recovery  
5 (Trondheim, Norway, 22-24 juin 1998), 109-117, 1998.

6 Tokunaga, K. T., Wan, J., Sutton, S. R., Transient film flow on rough fracture surfaces, *Water*  
7 *Resour. Res.*, 36(7), 1737-1746, 2000.

8 Tokunaga, K. T., Wan, J., Surface-zone flow along unsaturated rock fractures, *Water Resour.*  
9 *Res.*, 37(2), 287-296, 2001.

10 Toride N., Leij F. J., van Genuchten M. Th., The CXTFIT Code for Estimating Transport  
11 Parameters from Laboratory or Field Tracer Experiments, Version 2.0, Research Report  
12 No. 137, U.S. Salinity Laboratory, USDA, ARS, Riverside, CA, 121p, 1995.

13 Vanderkwaak, J. E., Forsyth, P. A., MacQuarrie, K. T. B., Sudicky, E. A., WatSolv sparse  
14 matrix iterative solver, User's guide for version 2.16, University of Waterloo, Waterloo,  
15 Ontario, Canada, 22p, 1997.

16 Vanderkwaak, J. E., Numerical simulation of flow and chemical transport in integrated  
17 surface-subsurface hydrologic systems, PhD Thesis, University of Waterloo, Department  
18 of Earth Sciences, 217p, 1999.

19 Van Genuchten, M. Th., A closed-form equation for predicting the hydraulic conductivity of  
20 unsaturated soils, *Soil Sci. Soc. Am. J.*, 44, 892-898, 1980.

21 Van Genuchten, M. Th., Wierenga, P. J., Mass transfer studies in sorbing porous media : I.  
22 Analytical solutions, *Soil Sci. Soc. Am. J.*, 40, 473-480, 1976.

23 Wang, J. S. Y., Narasimhan, T. N., Hydrologic mechanisms governing fluid flow in a  
24 partially saturated, fractured, porous medium, *Water Resour. Res.*, 21(12), 1861-1874,  
25 1985.

1 Yeh, G.-T., A lagrangian-eulerian method with zoomable hidden fine mesh approach to  
 2 solving advection-dispersion equations, Water Resour. Res., 26(6), 1133-1144, 1990.

3 Younger, P. L., Elliot, T., Chalk fracture system characteristics : implications for flow and  
 4 solute transport, Q. J. Eng. Geol., 28, S39-S50, 1995.

5 Zurmühl, T., Durner, W., Modeling transient water and solute transport in biporous soil,  
 6 Water Resour. Res., 32(4), 819-829, 1996.

7 **Annex 1: Mathematical expressions used to compute the derivatives of the**  
 8 **retention curve and the hydraulic conductivity curve**

9 As mentioned previously, the numerical solution of Richards' equation using Newton-  
 10 Raphson linearization technique needs computation of the derivatives of the retention curve  
 11 and the hydraulic conductivity curve according to the water content. For completeness, the  
 12 expressions are provided here after.

13 When fissures are fully desaturated and the matrix is partially saturated with water:

14 
$$\frac{d\theta}{dh} = \frac{-\alpha_M m_M (\theta_{s,M} - \theta_r)}{1 - m_M} \Theta_M^{1/m_M} (1 - \Theta_M^{1/m_M})^{m_M} \quad (\text{A.1})$$

15 
$$\frac{dk_r}{dh} = \frac{dk_r}{d\theta} \frac{d\theta}{dh} = \left( \frac{P}{\theta_s - \theta_r} S_e^{(P-1)} + 2S_e^P \frac{f_I(\theta)}{f(\theta_s)} \frac{df_I(\theta)}{d\theta} \right) \frac{d\theta}{dh} \quad (\text{A.2})$$

16 with :

17 
$$\Theta_M = \frac{\theta - \theta_r}{\theta_{s,M} - \theta_r}$$

18 
$$m_M = 1 - \frac{1}{n_M}$$

19 
$$\frac{df_I}{d\theta} = \alpha_M (1 - \Theta_M^{1/m_M})^{(m_M-1)} \Theta_M^{(1/m_M-1)}$$

20 When the matrix is fully saturated and fissures are partially saturated with water :

$$1 \quad \frac{d\theta}{dh} = \frac{-\alpha_F m_F (\theta_s - \theta_{r,F})}{1 - m_F} \Theta_F^{1/m_F} (1 - \Theta_F^{1/m_F})^{m_F} \quad (\text{A.3})$$

$$2 \quad \frac{dk_r}{dh} = \frac{dk_r}{d\theta} \frac{d\theta}{dh} = \left( \frac{P}{\theta_s - \theta_r} S_e^{(P-1)} + 2S_e^P \frac{(f_I(\theta_J) + f_{II}(\theta))}{f(\theta_s)} \frac{df_{II}(\theta)}{d\theta} \right) \frac{d\theta}{dh} \quad (\text{A.4})$$

3 with :

$$4 \quad \Theta_F = \frac{\theta - \theta_{r,F}}{\theta_s - \theta_{r,F}}$$

$$5 \quad m_F = 1 - \frac{1}{n_F}$$

$$6 \quad \frac{df_{II}}{d\theta} = \alpha_F (1 - \Theta_F^{1/m_F})^{(m_F-1)} \Theta_F^{(1/m_F-1)}$$

1 **Table captions**

2 Table 1. Description of tracer injection performed in the saturated (Pz CS) and in the  
3 unsaturated chalk (Pz CNS).

4

5 Table 2. Parameters defined for modelling the behaviour of unsaturated formations. For the  
6 loess and the flint conglomerate, the van Genuchten – Mualem model is used. For the chalk,  
7 the new functionality developed in section 1 is used.

8

9 Table 3. Fitted transport parameters for the chalk using the tracer test results.  $\alpha_L$  is the  
10 longitudinal dispersivity coefficient (L) appearing in the expression of the hydrodynamic  
11 dispersion tensor. (<sup>1</sup> Values of effective and immobile water porosity depend of the saturation  
12 degree of the rock. This may change from point to point and it is thus impossible to give  
13 values for these parameters).

14

## 1 **Figure captions**

2 Figure 1. Conceptual model for water recharge and solute transport mechanisms in the  
3 variably saturated chalk. Under intensive or artificial recharge conditions, water and  
4 contaminants migrate through the saturated fissures and the matrix acts as a buffer zone.  
5 Under “natural” recharge conditions, the fissure network is inactive; the migration of water  
6 and contaminants is restricted in the matrix.

7  
8 Figure 2. Illustration of the method used to construct the global retention curve relative to the  
9 fissured, dual porosity chalk (dashed line) from the superposition of two van Genuchten  
10 retention curves, one for the matrix (light grey curve), one for the fissure (dark grey curve)  
11 components. A continuously derivable condition is prescribed at the point  $(\theta_j, h_j)$ .

12  
13 Figure 3. Unsaturated characteristic curves obtained with  $\theta_{s,M} = 0.41$ ,  $\alpha_M = 0.099 \text{ m}^{-1}$ ,  
14  $n_M = 1.1$ ,  $\theta_s = 0.42$ ,  $\alpha_F = 10 \text{ m}^{-1}$ . The characteristic curve relative to the matrix and fissure  
15 components join at a suction value  $h_j = 0.4 \text{ m}$ , the continuity parameters obtained by solving  
16 equations (4a and b) being  $n_F = 3.11$  and  $\theta_{r,F} = 0.4088$ .

17  
18 Figure 4. Partitioning of water between the mobile and the immobile porosity as a function of  
19 the ratio between associated relative hydraulic conductivity values. At saturation (a), the  
20 mobile water is associated with the fracture porosity and the dual porosity effect is important;  
21 as the rock desaturates (b), the mobile water becomes associated with water in the matrix and  
22 the dual porosity effect is reduced.

23



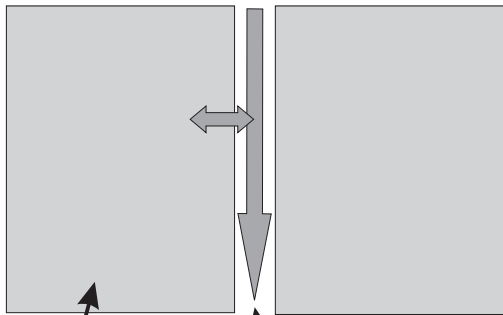
- 1 Figure 5. Breakthrough curves of tracers injected in Pz CS during the field tracer experiments
- 2 (line + symbols) and modelled with the SUFT3D code (dashed lines).
- 3
- 4 Figure 6. Breakthrough curves of the tracers injected in Pz CNS under (a) intense artificial
- 5 recharge conditions and (b) natural recharge conditions (lines + symbols) and modelling
- 6 results obtained with the SUFT3D code (lines).

Injection well	Pz CS phase 1	Pz CS phase 2	Pz CNS phase 1	Pz CNS phase 2
Injection conditions	Saturated zone	Saturated zone	Artificially intense recharge conditions	Natural recharge conditions
Pumping rate at PC (m <sup>3</sup> /h)	1.2	6.0	6.48	Permanent pumping at a rate varying between 3 and 6 m <sup>3</sup> /h
Tracer	eosin Y	naphtionate	KCl	I <sup>-</sup>
Injected mass (kg)	0.0053	0.0051	100	7.64 (10 of KI)
Tracer volume (m <sup>3</sup> )	0.010	0.010	0.3	0.03
Tracer injection duration (h)	0.031 (1min53s)	0.036 (2min11s)	0.88 (53 min)	0.1 (6 min)
Water flush volume (m <sup>3</sup> )	0.127	0.132	Constant recharge: 0.1 m <sup>3</sup> /h	No recharge
Water flush duration (h)	0.29 (17min20s)	0.22 (12min56s)		
Minimum transit time (h)	0.58 (35min)	0.25 (15min)	5.25	8184 (341 days)
Modal transit time (h)	1.42 (85min)	0.5 (30min)	11,5	15312 (638 days)?

Material		$\theta_r$ (-)	$\theta_{s,M}$	$\theta_{r,F}$	$\theta_s$ (-)	$\alpha_{VG}$ (m <sup>-1</sup> )	$n_{VG}$ (-)
Loess		0.01	-	-	0.445	3.0	1.14
Flint Conglomerate		0.065	-	-	0.41	1.0	1.89
Fissured chalk	Matrix	0.01	0.40	-	-	0.05	1.1
	Fissures	-	-	0.4088(*)	0.42	10	4.56 (*)
(*) parameters adjusted for guaranteeing continuity at the matching point defined on the matrix and fissure retention curves, for $h_j = 0.4$ m							

Injection well	Experimental conditions	Modelling approach	$\theta_m$ (-)	$\theta_{im}$ (-)	$\alpha_L$ (m)	$\alpha$ (s <sup>-1</sup> )
Pz CS	Saturated zone	CDPM	0.004	0.05	1.0	$2.0 \times 10^{-7}$
PsCNS	Intense recharge	CDPM	0.01	0.08	1.0	$2.3 \times 10^{-7}$
		DDPM	$c_{part} = 0.02^1$		1.0	$9.0 \times 10^{-8}$
	Natural infiltration	CDPM	0.10	0.15	1.0	$2.3 \times 10^{-7}$
		DDPM	no tracer arrival observed during the simulated period			

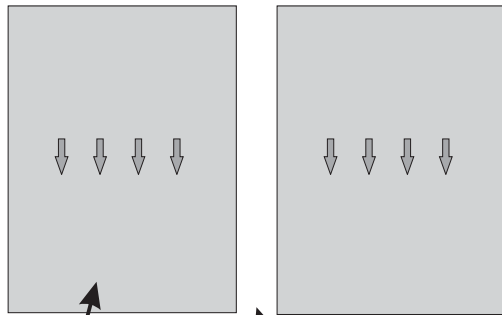
intensive rain / recharge



matrix blocks

active fissures

natural recharge



matrix blocks

inactive fissures

