

Assessment of Solvent Degradation within a Global Process Model of Post-Combustion CO₂ Capture

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Abstract

Solvent degradation may be a major drawback for the large-scale implementation of post-combustion CO₂ capture due to amine consumption and emission of degradation products. However, its influence on the process operations has rarely been studied. In the present work, a kinetics model describing solvent oxidative and thermal degradation has been developed based on own experimental results for the benchmark solvent, i.e. 30 wt% monoethanolamine (MEA) in water. This model has been included into a global Aspen Plus model of the CO₂ capture process. The selected process modelling approaches are described in the present work. Using the resulting simulation model, optimal operating conditions can be identified to minimize both the energy requirement and the solvent degradation in the process. This kind of process model assessing solvent degradation may contribute to the design of large-scale CO₂ capture plants to consider not only the process energy penalty, but also its environmental penalty. Indeed, both aspects are relevant for the large-scale deployment of the CO₂ capture technology.

Keywords: Post-combustion CO₂ capture, Solvent degradation, Process modelling.

1. Introduction

In a context of growing world energy demand and environmental concerns, CO₂ capture, re-use and storage is one of the most promising technologies to significantly and rapidly reduce greenhouse gas emissions. Among the different CO₂ capture methods, the post-combustion capture is the most mature one for a large-scale deployment. It consists in an absorption – regeneration loop in which CO₂ from the flue gas is absorbed in an amine solvent at temperatures varying between 40 and 60 °C. The process is usually designed so that the flue gas vented to the atmosphere after absorption contains 90 % less CO₂. The CO₂-loaded solvent is regenerated in a stripper at a higher temperature (between 100 and 140 °C, depending on the solvent). The produced CO₂ stream is almost pure and may be valorized (applications in food industry, enhanced oil recovery...) or stored underground. The present work studies the CO₂ capture in coal and natural gas fired power plants.

At the operational level, Svendsen et al. (2011) have identified two main drawbacks regarding post-combustion CO₂ capture. First, the high energy requirement for regenerating the amine solvent leads to a decrease by about 30 % of the plant efficiency. Then, the environmental impact of the CO₂ capture is affected by the emissions of amine solvent and solvent degradation products. Besides the emission of potentially harmful products, solvent degradation also impacts the process operating and capital costs. Indeed, the solvent absorption capacity decreases, its viscosity and its corrosivity increase, the vapour-liquid equilibrium is modified and foaming and fouling may appear

in the columns (Bedell, 2009). So far, the process energy penalty and the degradation of amine solvents have been studied separately and previously published models of the CO₂ capture process did not consider solvent degradation at all (e.g. Léonard and Heyen, 2011). Thus, the objective of the present work is to develop a model of the CO₂ capture process that assesses solvent degradation based on own experimental work, so that the influence of process operating conditions on solvent degradation may be better understood. As a result, solvent degradation could be considered more accurately during the design of CO₂ capture units, which is especially relevant for large-scale plants.

2. Experimental study

Two main solvent degradation mechanisms take place in the CO₂ capture process: oxidative degradation and thermal degradation with CO₂. However, few data were available to describe the influence of process operating conditions on solvent degradation. Because solvent degradation is a slow phenomenon, accelerated conditions are necessary to study it within a reasonable timeframe. Thus, an experimental test rig consisting in an agitated pressurized reactor with continuous gas feed has been designed for accelerating solvent degradation. Figure 1 shows this test rig which allows the degradation of solvents at temperatures up to 140 °C and pressures up to 20 bar, with flexible gas composition. Liquid degradation products are quantified using gas chromatography while the MEA content is determined by high performance liquid chromatography. Gaseous degradation products are quantified by Fourier transformed infra-red spectroscopy. The nitrogen mass balance of degradation experiments could be closed within 10 % and repeatability is demonstrated with a deviation lower than 5 %.

Various degradation experiments have been performed to study the influence of process operating variables on the degradation of 30 wt% MEA in water (benchmark solvent for CO₂ capture). Among other, the influence of the agitation rate, the temperature and the composition of the flue gas feed (varying concentrations in N₂, O₂ and CO₂) have been tested. Identified degradation pathways at lab-scale are observed to be similar to pathways observed in CO₂ capture pilot plants as evidenced in Figure 2 by the comparison of lab and industrial degraded solvent samples. The main products identified in Figure 2 have been listed in Table 1. Moreover, ammonia is the main degradation product identified in the gas phase. More details and results of the experimental degradation study are given in a previous work (Léonard, 2013).

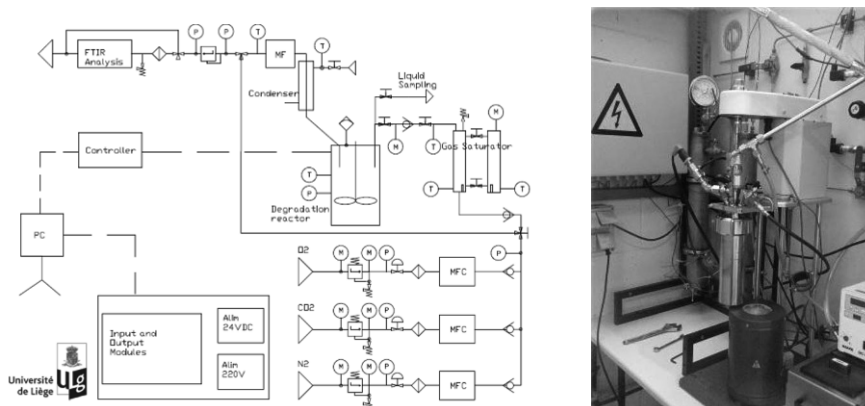


Figure 1. Degradation test rig designed and built at the University of Liège

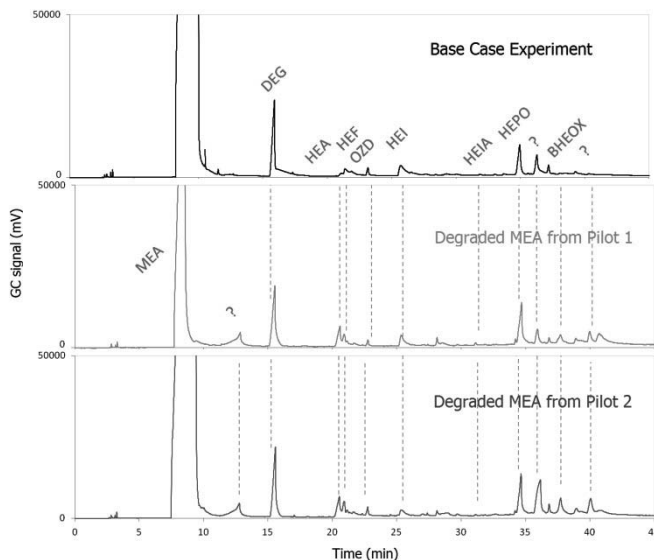


Figure 2. Comparison of the gas chromatography spectra between lab experiments (base case experiment) and degraded MEA samples from industrial pilot plants

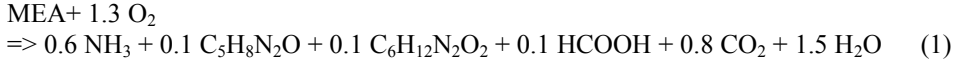
Table 1. Main identified products by gas chromatography of degraded MEA solutions

	Compound	Formula	Structure
MEA	Monoethanolamine	C ₂ H ₇ NO	<chem>HOCH2CH2NH2</chem>
DEG	Diethylene glycol (internal standard)	C ₄ H ₁₀ O ₃	<chem>HOCH2CH2OCH2CH2OH</chem>
HEA	<i>N</i> -(2-hydroxyethyl)acetamide	C ₄ H ₉ NO ₂	<chem>CC(=O)NCCO</chem>
HEF	<i>N</i> -(2-hydroxyethyl)formamide	C ₃ H ₇ NO ₂	<chem>C=O[NH]CCO</chem>
OZD	2-oxazolidinone	C ₃ H ₅ NO ₂	<chem>O=C1NCCO1</chem>
HEI	<i>N</i> -(2-hydroxyethyl)imidazole	C ₅ H ₈ N ₂ O	<chem>C1=CN=CN1CCO</chem>
HEIA	<i>N</i> -(2-hydroxyethyl)imidazolidinone	C ₅ H ₁₀ N ₂ O ₂	<chem>O=C1NCCN1CCO</chem>
HEPO	4-(2-hydroxyethyl)piperazine-2-one	C ₆ H ₁₂ N ₂ O ₂	<chem>O=C1NCCN1CCO</chem>
BHEOX	<i>N,N'</i> -Bis(2-hydroxyethyl)oxamide	C ₆ H ₁₂ N ₂ O ₄	<chem>O=C(NCCO)C(=O)NCCO</chem>

3. Model construction

Based on the results of the degradation study, a kinetics model of MEA degradation is proposed, assuming one empirical degradation reaction for each studied degradation pathway. Since the exact reaction mechanisms of MEA oxidative degradation are still unknown, the stoichiometric coefficients of degradation products are determined from the product distribution observed in the experimental study and normalized to the degradation of one mol MEA. Although not measured in the present work, formic acid

has been included to take the formation of heat stable salts into account since they have been identified in previous experimental studies (Sexton and Rochelle, 2009). The resulting apparent reaction of MEA with oxygen is represented in Equation (1) while its kinetics is given in Equation (2) according Arrhenius' equation. Regarding the thermal degradation of MEA with CO₂, degradation mechanisms are known and lead to Equation (3) with the associated kinetics from Equation (4). R is the universal gas constant (8.314 J/mol.K), T the temperature (K), and the MEA, O₂ and CO₂ concentrations are in mol/L. The values for the activation energy are given in J/mol.



$$-r_{\text{MEA, Oxidative}} = 5.35 \cdot 10^5 \cdot e^{-41,730/RT} \cdot [\text{O}_2]^{1.46} \quad (2)$$



$$-r_{\text{MEA, Thermal}} = 6.27 \cdot 10^{11} \cdot e^{-143,106/RT} \cdot [\text{CO}_2]^{0.9} \quad (4)$$

This kinetics model for MEA oxidative and thermal degradation with CO₂ has been included into a global rate-based process model developed in Aspen Plus (Léonard and Heyen, 2013). This model represents the pilot plant described by Knudsen et al. (2011), treating a flue gas flow rate of 5000 Nm³/h. In order to integrate the degradation into the CO₂ capture process, different process modelling assumptions have been considered. First, the degradation reactions have been included into a steady-state model since dynamic simulations are not adapted to describe small modifications over long time scales (several months). Then, the degradation reactions have been considered to take place in the absorption and stripping columns, which better reflects the actual process operating conditions. Degradation in other process equipment has been neglected. Moreover, the degradation model has been designed without any purge or MEA make-up stream. These streams were initially present in the model, but since the degradation rates are very low, accumulating degradation products did not prevent the simulation from converging even in the absence of purge and make-up streams. This implies that the calculation converges although the tear streams of the solvent loop are not rigorously in mass balance. As a consequence, degradation rates are quantified by the formation of degradation products in the mass transfer columns since column reactions are precisely calculated by Aspen Plus at the contrary of solvent loop concentrations. Finally, the degradation products appearing in the degradation reactions had to be specified in Aspen Plus. Component data for NH₃ and HCOOH have been retrieved from Aspen Plus databases. However, no data were available for the other organic degradation products, but they could be estimated by Aspen Plus based on their molecular structure. In a first approach, these components have been defined as non-volatile to facilitate the liquid-vapor equilibria calculations. Considering the high molecular weights of these components, this non-volatility assumption appears relevant.

4. Simulation results

The main observations reported for the base case configuration of the degradation model are listed in Table 2. Ammonia and HEIA production rates have been reported, since ammonia is the main product of MEA oxidative degradation and HEIA the main product of the MEA thermal degradation with CO₂. The model confirms that oxidative degradation in the absorber is the main cause for solvent loss in CO₂ capture plants. It also confirms that ammonia emissions at the absorber outlet are one of the major environmental issues of this technology (Mertens et al., 2013). The amount of degraded

MEA observed in the model is in the same order of magnitude compared with CO₂ capture plant results, although the predicted values are lower. Indeed, the total MEA loss equals 0.082 kg MEA/t CO₂ and recent pilot plant results (Moser et al., 2011) report a degradation rate of 0.284 kg MEA/t CO₂ in the absence of degradation inhibitors. This difference may be due to two assumptions done in first-approach, which are to neglect the presence of SO_x and NO_x contaminants in the flue gas, and to neglect the presence of dissolved metals in the solvent solution as well. Since all these components are known to increase the degradation rate (Sexton and Rochelle, 2009), they should be considered in further model developments.

Using the developed model, the influence of process operating variables on the energy and the environmental penalties of the CO₂ capture process can be quantified. The solvent flow rate and concentration, the stripper pressure and the oxygen concentration in the flue gas are varied. The largest energy savings are observed when increasing the stripper pressure and the solvent concentration, while the most significant increases of the MEA consumption are due to the increased oxygen concentration in the flue gas and to the higher solvent concentration. An example of such result is represented in Figure 3. It appears that increasing the MEA concentration from 30 wt% to 40 wt% reduces the reboiler heat duty by 4 % from 3.64 to 3.49 GJ/t CO₂. However, it also doubles the MEA loss from 0.082 up to 0.175 kg/t CO₂, so that concentrated MEA is not an advantageous solvent except if degradation inhibitors are added to the solvent solution. Finally, two process improvements are tested (absorber intercooling and lean vapour compression) that both reduce the total process energy requirement with no significant influence on solvent degradation. Combining these two improvements with optimal operating conditions leads to a reduction of the process energy requirement by 12 % at equivalent solvent consumption rate. The corresponding energy requirement of the CO₂ capture process is evaluated to 3.1 GJ/t CO₂, with a resulting MEA consumption of 78 g MEA/t CO₂.

Table 2. Degradation and emission results for the degradation model

Parameter	Unit	Absorber	Stripper	Total
MEA degradation	kg/t CO ₂	8.2e-2	1.4e-5	8.2e-2
NH ₃ formation	kg/t CO ₂	1.4e-2	8.4e-7	1.4e-2
HEIA formation	kg/t CO ₂	1.1e-5	1.1e-5	2.2e-5
MEA emission	kg/t CO ₂	8.7e-4	9.4e-9	8.7e-4
NH ₃ emission	kg/t CO ₂	9.5e-3	3.0e-3	1.3e-2
Top stage liquid temperature	°C	57.4	96.6	-
Bottom stage liquid temperature	°C	51.1	115.6	-
Top stage vapor O ₂ content	mol%	6.3	9.6e-3	-
Bottom stage vapor O ₂ content	mol%	6.1	7.0e-14	-

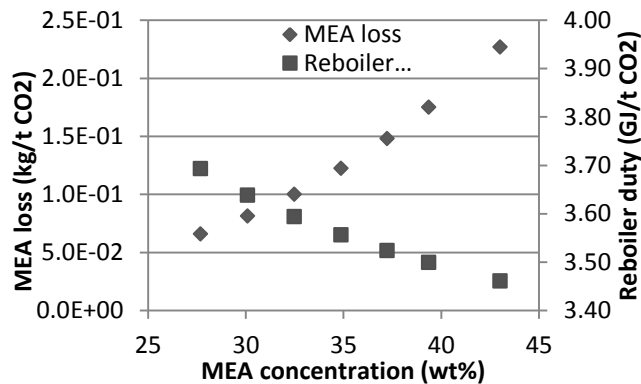


Figure 3. Influence of the MEA concentration on the process energy and environmental penalties

5. Conclusions

The present work may be considered as a first step for assessing both solvent degradation and energy requirement in CO₂ capture plants. This methodology that has been developed for the case of MEA in CO₂ capture conditions may be extended to alternative promising solvents, like piperazine for instance. This kind of model could and should be used for the capture plant design to consider not only the process energy penalty, but also its environmental penalty. Indeed, both aspects are particularly relevant in case of large-scale deployment of the CO₂ capture technology. Further modelling perspectives may consider the catalytic effect of dissolved metals on solvent oxidative degradation and the presence of SO_x and NO_x contaminants in the power plant flue gas.

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